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Title:

**Effects of nano carbon black and single-layer graphene oxide on settlement,  
survival and swimming behaviour of *Amphibalanus amphitrite* larvae**

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26 **Abstract**

27

28 The effects of two carbon-based nanomaterials, nano-sized carbon black (nCB) and single-  
29 layer graphene oxide (GO), on settlement of *Amphibalanus amphitrite* (Cirripedia, Crustacea)  
30 cypris larvae (cyprids) were assessed after 24, 48 and 72 h of exposure. Additionally, the  
31 effects of these nanomaterials on the mortality and swimming behaviour of the nauplius  
32 larvae (nauplii) of the same organism were determined after 24 and 48 h. The data indicate  
33 that nCB is more effective as a potential antisetlement agent than single-layer GO; moreover,  
34 nCB did not show any adverse effects on the larvae. The swimming behaviour of II stage  
35 nauplii of *A. amphitrite* exposed to a suspension of nCB was inhibited only at very high nCB  
36 concentrations ( $\geq 0.5$  mg/mL). Single-layer GO, on the contrary, showed lower antisetlement  
37 effects and was more active in altering the survival and inhibiting the swimming behaviour of  
38 the nauplii. An indication of the toxic or non-toxic mechanisms of the antisetlement  
39 properties of both of these nanomaterials is provided by the reversibility of the antisetlement  
40 activity. In conclusion, we propose nCB as an innovative antifouling nanomaterial that shows  
41 low toxicity towards the model organism (crustaceans) used in this study.

42

43 **Keywords:** *Amphibalanus amphitrite*; antifouling; nanotoxicity, nanomaterials; nano carbon  
44 black; single-layer graphene oxide.

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46 **Word count:** 3,041.

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## 48 **1. Introduction**

49

50 From the initial adsorption of organic molecules, through the colonisation by microorganisms,  
51 to the development of complex sessile assemblages, biofouling affects most man-made  
52 surfaces, such as ship hulls, oilrigs, mariculture cages, pipelines, heat exchangers and  
53 seawater intakes in general, which results in significant economic costs [1]. Biofouling is a  
54 problem that can affect any artificial device in contact with water, so different antifouling  
55 strategies are being designed to prevent the attachment and growth of microfouling and  
56 macrofouling organisms on man-made submerged surfaces.

57 Most of the commonly used antifouling paints contain biocides. The key property of a  
58 good antifouling biocide with respect to the environment is that it is effective in preventing  
59 fouling without causing persistent adverse environmental effects. The effects and behaviour  
60 of biocides used in antifouling paints have been extensively studied [2,3]. Currently, a wide  
61 range of chemicals are used as antifouling biocides, which are governed by different  
62 regulations, and which depend on the legislation in each country. In the search for innovative  
63 and less harmful antifouling materials, several metal nanomaterials are being tested, and these  
64 have shown promising results to date for the prevention of biofilm formation [4-8]. Studies on  
65 the antisettlement effects of carbon-based nanomaterials have also been reported recently;  
66 e.g., synthetic carbon nanotubes have been shown to be very efficient for the prevention of the  
67 formation of microbial biofilms [9], as well as against fouling by zooplanktonic and  
68 phytoplanktonic larvae [10]. It was also recently shown that graphene oxide (GO) can be used  
69 for the prevention of protein adsorption in microfluidic systems [11]. To our knowledge, no  
70 data on antisettlement effects of other carbon nanomaterials towards microfouling and  
71 macrofouling have been reported. Furthermore, although some different nanomaterials have  
72 shown good potential as antifouling agents, present knowledge relating to their adverse

73 effects in aquatic environments is very limited [12-14]. It is therefore very important to gain  
74 more data about the effects of these materials before they are used globally, and before  
75 different ecosystems are exposed to them.

76 The aim of the present study was therefore to evaluate the effects of two different  
77 carbon nanomaterials, nano-sized carbon black (nCB) and single-layer GO, on the settlement  
78 of *Amphibalanus amphitrite* cypris larvae (cyprids), combined with an assessment of their  
79 toxicity towards the naupliar larval stage (nauplii) of the same organism. The barnacle, *A.*  
80 *amphitrite*, is a cosmopolitan cirriped crustacean that represents one of the main species of  
81 biofouling organisms. Indeed, it is reared and used mainly as a model organism for efficacy  
82 and toxicity testing of antifouling products. The ease of obtaining and rearing its larval stages  
83 has extended its use also as a model for ecotoxicological studies [15,16].

84 Nano-sized carbon black is a form of amorphous carbon that has a high surface area-  
85 to-volume ratio. It is a product of the incomplete combustion of fossil fuels and vegetation  
86 [17], and it is used in rubber production, as a black pigment in printing inks, as electrodes in  
87 batteries, and in leather production [18]. Graphene oxide is a two-dimensional nanomaterial  
88 that has single, double or a few ( $\leq 10$ ) layers of carbon atoms arranged in six-membered rings  
89 [19]. It has unique electrical, chemical and mechanical properties, such as a large surface area,  
90 excellent conductivity, high mechanical strength, and ease of functionalisation and mass  
91 production [20]. For this reason, GO is one of the most exciting materials under investigation.  
92 In the present study, we used single-layer GO; e.g., a graphene sheet with additional oxygen  
93 functional groups attached to its edges (carboxylic groups) or to its basal plane (phenol  
94 hydroxyl and epoxide groups) [21].

95

## 96 **2. Materials and methods**

97

## 98 **2.1. Materials**

99 The nCB powder was provided by PlasmaChem GmbH (Berlin, Germany), with the average  
100 size of the primary particles given as 13 nm. Dry flakes of single-layer GO were purchased  
101 from Graphene Supermarket (USA). According to the producer, this GO is composed of  
102 carbon (79%) and oxygen (20%). The flake size was assessed as between 0.5  $\mu\text{m}$  and 5  $\mu\text{m}$ .  
103 Stock suspensions were prepared of 1 mg/mL nCB or single-layer GO, in 0.22- $\mu\text{m}$ -filtered  
104 natural sea water (FNSW); these were sonicated for 15 min using a 50% on/off cycle. These  
105 stock suspensions were kept on ice during the preparation. Further characterisation of the  
106 secondary characteristics of these nanomaterials (e.g., dynamic light scattering,  $\zeta$ -potential)  
107 was not possible as they did not form stable suspensions in FNSW.

108

## 109 **2.2. Methods**

110

### 111 **2.2.1. Settlement test**

112 The cypris larval phase of *A. amphitrite* is competent for metamorphosis and settlement, and  
113 these cyprids were obtained from laboratory cultures of a brood stock. Before being used in  
114 settlement assays, the larvae were filtered and maintained in FNSW at 4 °C for 3 days [22].  
115 After 3 days, the settlement tests were performed by adding 15-20 cyprids to each well of a  
116 24-well polystyrene plate that contained 2 mL of different concentrations of the suspended  
117 nCB (0.01, 0.1, 0.2, 0.3 mg/mL) or single-layer GO (0.01, 0.1, 0.5 mg/mL). The plates were  
118 stored at 28 °C for 72 h, under 16-h:8-h light/dark cycles. All of the settlement tests were  
119 performed in triplicate. After 24, 48 and 72 h, the number of settled and non-settled cyprids  
120 was examined under a stereomicroscope. The concentration of the nanomaterial that resulted  
121 in 50% settlement inhibition of these exposed organisms ( $\text{EC}_{50[\text{cyprids}]}$ ) was calculated at 72 h.  
122 Additionally, after 72 h, the settlement recovery of the cyprids was tested: the unsettled larvae

123 were collected from the wells at 72 h, and they were rinsed and transferred into new plates  
124 with clean FNSW. After an additional 24 h at 28 °C, the settlement percentage of these  
125 transferred larvae was determined again.

126

### 127 **2.2.2. Naupliar toxicity test**

128 The nauplii were obtained from laboratory cultures of a brood stock of *A. amphitrite*. This  
129 toxicity test was performed by adding 15 to 25 II stage nauplii to each well of a 24-well  
130 polystyrene plate that contained 1 mL of different concentrations of the suspended nCB (0.1,  
131 0.5, 1, 5 mg/mL) or single-layer GO (0.001, 0.01, 0.1, 0.5, 0.75, 1 mg/mL). The plates were  
132 stored at 20 °C for 24 h and 48 h, under 16-h:8-h light/dark cycles. All of the toxicity tests  
133 were performed in triplicate. After 24 h and 48 h, the nauplii were collected from the wells,  
134 rinsed, and transferred into new plates with clean FNSW. The number of dead larvae was  
135 determined by examination under a stereomicroscope. The concentration of the nanomaterial  
136 that resulted in 50% mortality of these exposed organisms ( $LC_{50[nauplii]}$ ) was calculated after  
137 24 h and 48 h of exposure.

138

### 139 **2.2.3. Naupliar swimming speed alteration test**

140 The naupliar swimming speed alteration test was performed according to [15, 23]. Briefly, 15  
141 to 25 II stage nauplii were exposed to the nanomaterials as described in paragraph 2.2.2. After  
142 24 h and 48 h, the swimming of the nauplii was recorded using a Swimming Behavioural  
143 Recorder experimental set-up. A Swimming Behavioural Recorder consists of a video camera  
144 with a macro-objective that records the paths of a sample of larvae swimming. The apparatus  
145 is caged inside a black box (60 × 60 × 100 cm) to exclude external sources of light, and the  
146 recording chamber is monitored under infrared light. The nauplii were dark-adapted for 2 min  
147 before starting the video recording (a time that was fixed in preliminary tests, to allow their

148 steady speed and uniform spatial distribution). The swimming behaviour was digitally  
149 recorded for about 3 s at 25 frames/s, and the images were analysed using advanced image  
150 processing software (the Swimming Behavioural Recorder system developed by e-magine IT,  
151 Genova, Italy). This analysis provided reconstructions of individual nauplius path-tracks, and  
152 measurements of the average swimming speed (mm/s) for each sample (of 15-20 individuals).  
153 The data are finally referred to as the swimming induction or inhibition, which was  
154 normalised to the average swimming speed of the control (where  $S$  is the average swimming  
155 speed), and Inhibition (%) =  $((S_{\text{Treated}} - S_{\text{Control}}) / S_{\text{Control}}) \times 100$ ). After the image analysis, the  
156 presence/ absence of individual tracks of each sub-sample of larvae (control and treated  
157 groups) provided the estimation of the swimming induction or inhibition.

158

#### 159 2.2.4. Statistical analysis

160 The  $EC_{50[\text{cyprids}]}$ , as the settlement inhibition of the cyprids after 72 h, the  $EC_{50[\text{nauplii}]}$ , as the  
161 swimming speed effects on the nauplii after 24 h and 48 h, and the  $LC_{50[\text{nauplii}]}$  mortality, and  
162 their related 95% confident limits after 24 h and 48 h, were calculated using trimmed  
163 Spearman–Karber analysis [24].

164

### 165 3. Results

166

#### 167 3.1. Settlement test

168 The results of the settlement inhibition test are shown in Figure 1. [insert Figure 1 near  
169 here] Both of these carbon nanomaterials showed dose-dependent antisettlement effects on  
170 the *A. amphitire* cypris larvae. The antisettlement effect of nCB on these cyprids was  
171 prominent at all exposure times (24 h, 48 h, 72 h) at nanomaterial suspensions ranging from  
172 0.01 mg/mL to 0.3 mg/mL. After 24 h of exposure to 0.01 mg/mL nCB, complete settlement

173 inhibition was observed. At 48 h and 72 h, the inhibition effect of nCB decreased. In the  
174 concentration range from 0.1 mg/mL to 0.3 mg/mL, nCB completely inhibited the settlement  
175 of the cyprids at all exposition times. The calculated  $EC_{50[\text{cyprids}]}$  (after 72 h) was 0.03 mg/mL.  
176 Single-layer GO had slightly lower effects in terms of settlement inhibition. The data show a  
177 more gradual decrease in the settlement inhibition at all of the exposure times (24 h, 48 h, 72  
178 h) and complete inhibition at 0.5 mg/mL of single-layer GO only after 24 h and 48 h, while  
179 after 72 h, the settlement percentage was around 40%. The calculated  $EC_{50[\text{cyprids}]}$  (after 72 h)  
180 was 0.15 mg/mL.

181         Interestingly, in both cases, the inhibition of settlement showed a reversible effect after  
182 rinsing the cyprids and exposing them for 24 h to FNSW without added nCB or single-layer  
183 GO. In the case of nCB, the reversibility was tested at all of the concentrations used, while  
184 single-layer GO was assayed only at the highest concentration of 0.5 mg/mL, due to the very  
185 high cost of the graphene particles. In both cases, at the highest tested concentration (0.3  
186 mg/mL nCB, 0.5 mg/mL single-layer GO), the recovery from the settlement activity was  
187 complete.

188

### 189 **3.2. Naupliar toxicity and swimming test**

190 The stage II nauplii of *A. amphitrite* behaved in different ways when they were exposed to the  
191 two nanomaterials tested.

192         The data showing the adverse effects of exposure of the nauplii of *A. amphitrite* to  
193 nCB for 24 h and 48 h are illustrated in Figure 2. [insert Figure 2 near here] Here, following  
194 small increases in the swimming behaviour at both 24 h and 48 h, higher concentrations of  
195 nCB induced increasing inhibition of the swimming behaviour. Also, although after 24 h of  
196 nCB there was no mortality of the nauplii at any of the nCB concentrations (Fig. 2A), after  
197 the 48-h exposure time, there was naupliar mortality at 5 mg/mL Ncb (Fig. 2B).

198 Similar effects, although at lower concentrations of nanomaterial, were observed after  
199 exposure of the nauplii to single-layer GO for 24 h (Fig. 3A) and 48 h (Fig. 3B). **[insert**  
200 **Figure 3 near here]** These indicate that increasing concentrations of single-layer GO and  
201 increasing exposure times lead to decreases in their swimming speed and to increases in their  
202 mortality. Again, as with nCB, although more pronounced, at the lower single-layer GO  
203 concentrations, there was induction of an increase in the swimming behaviour. This  
204 phenomenon was observed at up to 0.5 mg/mL and 0.01 mg/mL single-layer GO after 24-h  
205 and 48-h exposure, respectively. Higher concentrations of single-layer GO led to increased  
206 mortality and decreased swimming speeds, both of which occurred in a concentration-  
207 dependent manner particularly after the 48-h exposure time.

208 Table 1 gives the calculated  $LC_{50[nauplii]}$ , the swimming speed inhibition  $EC_{50[nauplii]}$  and  
209 settlement inhibition  $EC_{50[cyprids]}$  after 24 h and 48 h. The  $LC_{50[nauplii]}$  and swimming speed  
210 inhibition  $EC_{50[nauplii]}$  were not calculable after 24 h of nCB exposure, conversely to the  
211 single-layer GO exposure. **[insert Table 1 near here]**

212 The therapeutic ratio is the ratio between the amount of a therapeutic agent that  
213 provides a benefit and the amount that causes harm (i.e., death or toxicity). Thus, here the  
214 therapeutic ratio is defined as  $LC_{50[nauplii]}/$  settlement inhibition  $EC_{50[cyprids]}$ , and as indicated in  
215 Table 1, the calculated therapeutic ratios are 61.3 for nCB and 3.7 for single-layer GO.

216

#### 217 **4. Discussion**

218

219 The carbon-based nanomaterials used in this study show antifouling effects on *A. amphitrite*  
220 larvae, causing 50% inhibition of settlement ( $EC_{50[cyprids]}$ ) at concentrations of 0.03 g/mL nCB  
221 and 0.15 g/mL single-layer GO. These calculated  $EC_{50}$  values are higher with respect to the  
222 antifouling booster biocides, such as zinc-pyrithione, with an  $EC_{50}$  of  $2 \times 10^{-5}$  mg/mL [25],

223 although particularly in the case of nCB, the settlement inhibition  $EC_{50[\text{cyprids}]}$  is comparable to  
224 the  $EC_{50}$  obtained for antifouling compounds of natural origin [26].

225 Both nCB and single-layer GO induced increases in naupliar mortality, and decreases  
226 in the swimming speed of the nauplii in concentration-dependent manners. In addition, the  
227 nauplii exposed to both of these carbon-based nanomaterials show hormetic dose-response  
228 curves at 24 h and 48 h, whereby there is a stimulatory effect at low concentrations, and an  
229 inhibitory effect at higher concentrations.

230 The antisettlement activities induced by nCB and single-layer GO appear not to be due  
231 to the toxicity of these selected nanomaterials, since the naupliar mortality at concentrations  
232 that completely inhibited settlement was very low (*ca.* 2%) for nCB and about 20% for single-  
233 layer GO. To define this further, we also calculated the therapeutic ratio (or index), which is  
234 generally given as the ratio between the amount of a therapeutic agent that provides a benefit  
235 and the amount that causes harm (i.e., death or toxicity). Here, the therapeutic ratio  
236 ( $LC_{50[\text{nauplii}]} / \text{settlement inhibition } EC_{50[\text{cyprids}]}$ ) is thus an index used to estimate whether  
237 settlement inhibition is due to toxicity or other mechanisms, as has also been calculated across  
238 a number of similar studies [22, 27, 28]. A review from Qian et al. [29] on antifouling agents  
239 underlined how the therapeutic ratio is commonly used as a yardstick of the potential  
240 antifouling activity of a compound. In this review, it was suggested that a compound with a  
241 therapeutic ratio  $>15$  can be considered as a non-toxic antifouling compound, even if a much  
242 higher therapeutic ratio is recommended when selecting candidate compounds. Furthermore,  
243 Qian et al. [29] suggested that only a therapeutic ratio  $>50$  and an  $EC_{50} <5$  mg/L for the  
244 inhibition of both hard and soft fouling organisms should allow the antifouling compound to  
245 be considered as environmentally friendly. Following these recommendations, only nCB  
246 shows both of these requisites; conversely, single-layer GO does not satisfy these criteria.  
247 Moreover, an indication of the non-toxicity of the antifouling mechanism is provided by

248 consideration of the reversibility of the antisettlement activity: after the 72-h incubation with  
249 nCB or single-layer GO, when the cyprids were rinsed and transferred to clean FNSW, they  
250 settled again, thus completely regaining their settlement ability after 24 h, when the  
251 percentage of settlement was comparable to that obtained in the control settlement assay.

252         The exact mechanism of the antisettlement activity of the nanomaterials used in the  
253 present study remains to be clarified, and will be one of the main subjects of our future  
254 studies. The antisettlement effects of both of these nanomaterials used might be due to a  
255 simple mechanical interaction of these nanomaterials with the *A. amphitrite* cypris larvae, as  
256 both of these nanomaterials were not stable in FNSW. Such nanomaterials might alter the  
257 points of contacts of the cyprid antennular discs with the surface during the exploration phase  
258 before the settlement, which might therefore inhibit the whole process. In other words, the  
259 nanomaterials that precipitated to the bottom of the each well of the 24-well polystyrene  
260 multiwell plates (and on other submerged surfaces in natural environments) might alter the  
261 roughness of the surface, thus decreasing drastically the possibility of larval attachment. The  
262 effect of a nanostructured surface on the settlement of various representatives of fouling  
263 marine macroorganisms has also been shown recently by Scardino et al. [30]. In their study,  
264 superhydrophobic coatings with different physical architectures were prepared by bonding  
265 fumed silica nanomaterials to polysiloxane and its derivatives. Scardino et al. [30] showed a  
266 clear correlation between the settlement inhibition and the presence of “nanoroughness”.

267         Finally, very little or nothing is known about the adverse effects of nCB or single-  
268 layer GO on marine organisms. Rosenkranz et al. [31] reported that the LC<sub>50</sub> of nCB to  
269 freshwater *Daphnia magna* after 96 h is 0.4 mg/L. Our data show lower effectiveness of nCB  
270 versus *A. amphitrite* after 24 h and 48 h, where the LC<sub>50[nauplii]</sub> was 1.84 mg/mL, while the  
271 single-layer GO showed a lower LC<sub>50[nauplii]</sub> (0.56 mg/mL). The secondary characteristics of  
272 these nanomaterials, such as those expressed by dynamic light scattering and ζ-potential,

273 might contribute to an understanding their bioactivity. Thus, their detailed particle  
274 characterisation is needed to have a better view towards an understanding the results of the  
275 present study.

276

## 277 **5. Conclusions**

278

279 Two carbon-based nanomaterials were tested here to determine their effects on the settlement,  
280 mortality and swimming behaviour of *A. amphitrite* larvae, to investigate the possibility of  
281 finding new compounds for antifouling applications that show low toxicity. Our data indicate  
282 that nCB has less pronounced effects on the behaviour and survival of *A. amphitrite* larvae  
283 while having a considerable antisettlement effect. Single-layer GO shows more pronounced  
284 adverse effects, together with a lower antisettlement effects compared to nCB. Both of these  
285 nanomaterials can still be considered as potentially suitable as antifouling nanomaterials that  
286 might be incorporated into innovative paint formulations, although further investigations of  
287 their secondary characteristics, their toxicities towards other classes of organisms, and their  
288 behaviour in marine environments are required.

289

290

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388

389 **Table 1.** Calculated LC<sub>50</sub>[nauplii], swimming speed inhibition EC<sub>50</sub>[nauplii], settlement inhibition EC<sub>50</sub>[cyprids], and therapeutic ratio from the present  
 390 study.

391

Nanomaterial	LC <sub>50</sub> (nauplii)		Swimming speed inhibition EC <sub>50</sub> (nauplii)		Settlement inhibition EC <sub>50</sub> (cyprids)	Therapeutic ratio*
	(24 h)	(48 h)	(24 h)	(48 h)	(72 h)	
	nCB	>1 (1.66-2.03)	1.84	nc	0.48 (0.42-0.55)	
Single-layer GO	<1 (nc)	0.56 (0.51-0.61)	0.84 (0.82-0.86)	0.31 (0.27-0.35)	0.15 (nc)	3.7

392

393 nCB, nano-sized carbon black

394 GO, graphene oxide

395 Data are means (95% confident limits)

396 nc, not calculable; SSA, swimming speed alteration.

397 \*, LC<sub>50</sub>[nauplii] (48 h)/ settlement inhibition EC<sub>50</sub>[cyprids]

398

399 **Figure captions**

400

401 **Figure 1.** Settlement of *A. amphitrite* cypris larvae after 24, 48 and 72 h with different  
402 concentrations of nCB (**A**) and single-layer GO (**B**). Data are means  $\pm$ standard error (n = 3).  
403 The settlement reversibility (after 96 hours) is also shown for these cyprids (nCB, at all tested  
404 concentrations; single-layer GO, only at 0.5 mg/mL).

405

406 **Figure 2.** Effect of nCB on mortality (black bars) and swimming speed (white bars) of *A.*  
407 *amphitrite* nauplii after 24 h (**A**) and 48 h (**B**). Data are means  $\pm$ standard error (n = 3).

408

409 **Figure 3.** Effects of single-layer GO on mortality (black bars) and swimming speed (white  
410 bars) of *A. amphitrite* nauplii after 24 h (**A**) and 48 h (**B**). Data are means  $\pm$ standard error (n =  
411 3).

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413